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The capacities of institutions for the integration of ecosystem services in coastal strategic planning: The case of Jiaozhou Bay Li, R., Li, Y., van den Brink, M. and Woltjer, J.

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5 1. Introduction

Coastal areas are difficult to manage because they involve dynamic natural systems that are 6 increasingly under pressure from expanding socio-economic systems (Turner, 2000). One 7 central challenge for coastal management and planning in practice is to develop innovative 8 approaches for managing diverse human uses of ecosystems through a range of activities 9 10 (Lester et al., 2010). To meet this challenge, an ES approach has been increasingly adopted in ecosystem-based coastal management, marine spatial planning and strategic environmental 11 assessment (e.g., Partidario & Gomes, 2013; Böhnke-Henrichs et al., 2013). The concept of 12 13 ES provides a lens through which we can understand the relationships between humans and 14 natural systems. Specifically, this notion helps us assess how these services benefit humanity 15 and how human actions generally impact ecosystems and the delivered ESs (MA, 2005; 16 Carpenter et al., 2009). The Millennium Ecosystem Assessment (2005) developed four 17 broadly employed ES categories to help understand the above question: provisioning, regulating, cultural and supporting services. 18

19 A key difficulty in integrating these services into natural resource management and planning 20 is their complex and dynamic interrelationships in terms of trade-offs and synergies. Tradeoffs arise when the attempt to optimize a single service leads to reductions or losses of other 21 services (Holling & Meffe, 1996). A typical example would be a situation where offshore 22 23 wind farm development enhances energy production but simultaneously has negative impacts on biodiversity (Busch et al., 2011). ES synergies often arise when multiple services are 24 enhanced simultaneously (Raudsepp-Hearne et al., 2010). For instance, marine protection 25 areas maintain habitats while also producing important benefits for certain fish (Shen et al., 26 2011). These interrelationships usually emerge when several services respond to a driver 27 28 modified by human management or due to the interplay between ESs (Bennett et al., 2009). It has been argued that making these interrelationships explicit is a key informational need for 29 policy-making. More clarity on these interrelationships may reduce the risk of negative trade-30 offs and enhance potential win-win scenarios (Bennett et al., 2009; Lester et al., 2013; Kelble 31 32 et al., 2013).

Consequently, there has been increasing interest in developing decision-making approaches
based on analyzing ES interrelationships (Butler et al., 2013). Scholars typically use
economic valuation, geospatial information and multiple stakeholders' objectives to quantify

36 ES values or geographical clusters across landscapes and seascapes. Current approaches for

37 measuring ES trade-offs and/or synergies can be broadly grouped into four main approaches:

38 mapping (e.g. Costanza et al., 1998; Martínez-Harms & Balvanera, 2012; Crossman et al.,

39 2013), modeling (e.g. Swallow et al., 2009; Chisholm, 2010), social-survey analysis (e.g.

40 Hauck et al., 2013; Potts et al., 2014), and content analysis (Piwowarczyk et al., 2013;

41 Wilkinson et al., 2013). A large number of recent studies have used hybrid methods of

42 mapping and modeling (e.g. InVEST and ARIES; Nelson et al., 2009; Villa et al., 2009), or

43 mapping and social-survey analysis (e.g. SolVES; Sherrouse et al., 2011). Such approaches

44 have also been employed in the field of coastal and marine management to ascertain the

45 influence of diverse activities on key ESs. Examples involve reclamation, fisheries,

46 aquaculture, offshore wind farming, special marine protected areas, and wetland

47 developments that impact varying ESs (e.g., Brown et al., 2001; Martinet & Blanchard, 2009;

48 Busch et al., 2011).

The studies mentioned earlier mainly show people's general preferences for different service 49 50 categories: people tend to be less appreciative of regulating services and supporting services that create high-value provisioning and cultural services (Carpenter et al., 2006; Rodríguez et 51 52 al., 2006). In fact, scientists have emphasized the critical and vulnerable roles of regulating and supporting services (e.g., water purification, climate and flooding regulation, wetland 53 habitat and biodiversity) in various ES interrelationships. However, both these ES categories 54 55 are easily threatened by investment primarily in provisioning services (Bennett et al., 2009). 56 In addition, the studies mentioned earlier also suggest that close interrelationships among ES 57 are not well-articulated or handled in current coastal policy-making or planning (Halpern et 58 al., 2008). It is particularly true in coastal strategic planning, which generally refers to a 59 framework for arranging coastal and marine spatial use and organizing human activities to achieve economic and social benefits while sustaining ecosystem health, function and 60 services. Current coastal strategic planning has been unable to make ES trade-offs and 61 synergies explicit, especially when indirect effects make the identification and assessment of 62 the interplay of ESs more complex than simple cause-effect mechanisms (Halpern et al., 63 2008). Moreover, when either the spatial scale (in-site or off-site effects of interrelationships) 64 or the temporal scale (short-term or long-term effects) increases, ES interlinks could become 65 more uncertain and difficult to manage (Rodríguez et al., 2006). This would restrict the ability 66 of policy and planning to be more sustainable and adaptive. 67

68 New approaches to coastal strategic planning are increasingly important to addressing the

69 issues of sustainable and adaptive coastal and sea use. Although current research on

70 approaches for assessing ES interrelationships has contributed to decision-making in a variety

71 of ways, there are two main limitations. First, no attempt has been made to systematically

2

72 clarify the integration of ES interplay from coastal strategic plans in practice. There has been

73 a lack of attention to understand causal ES interrelationships embedded in actual coastal

74 policies. The second limitation is that most approaches do not handle a wide scope of drivers

75 and related ESs, and often lack an understanding of institutional contexts that determine

76 which specific driving forces, ESs and their interrelationships may be taken into account.

77 Therefore, the specific objective of this research is to propose a four-step method to assess a 78 broad range of drivers and ES interrelationships included in coastal strategic planning, based 79 on a more causal analyzing mechanism. In this way, this paper aims to clarify ES 80 interrelationships formulated in policy language, and it aims to provide insights into complex aspects of the coastal environment, from non-academic and strategic-policy points of view. 81 82 Such views may enable strategic planning to be more adaptive and sustainable in coastal areas where the integration of ESs for realizing ecosystem-based coastal management and planning 83 84 is in an early stage of development. Jiaozhou Bay in China is used as an illustrative case. The following section will introduce the background of this case. Next, we will explain our four-85 step method. After reporting the findings by applying the method, we will analyze the results, 86 discuss institutional implications for the consideration of the drivers and ES interrelationships 87 and, finally, reflect on our method's strengths and its implications. 88

89 2. The case study: Jiaozhou Bay in China

90 Jiaozhou Bay is a semi-enclosed and fan-shaped natural bay located on the southern coast of

91 Shandong Peninsula in East China (Fig. 1). In 2012, it covered an area of 343.5 km² and its

92 coastline measured 206.8 km. Several rivers feed into this bay, of which the largest is the

93 Dagu River. Seven districts and five county-level cities (all belonging to Qingdao City)

94 surround the bay, with a total population of 8.71 million.

95 We chose Jiaozhou Bay as a case study for several reasons. First, the development of the 96 whole urban area around the bay essentially depends on a large range of ESs provided by the 97 bay, such as aquaculture, fisheries, transportation, sea sports, tourism and large wetland maintenance (Zhao et al., 2005). A great deal of research on the ecological, physical, 98 99 chemical environment of Jiaozhou Bay has been extensively conducted (e.g. Shen 2001; Liu et al., 2004; Gao et al., 2014). The rich diversity in coastal and marine services and 100 101 understandings of the ecosystem yield useful ES information for strategic planning. The 102 second consideration concerns the importance of identifying how coastal activities may be 103 considered as drivers in the formation of ES interrelationships in strategic planning. Coastal 104 areas where fast-paced and long-term development takes place are more likely to provide answers, since intensive anthropogenic pressures result in different conflicts about ESs. This 105

106 is particularly the case in Qingdao – a leading coastal city in China and an economic center in 107 Shandong Province – whose extractive, industrial, commercial, recreational and emerging ocean uses have shrunk the area of Jiaozhou Bay by 173 km² (nearly one-third) over the past 108 109 45 years as a result of extremely rapid resource development (Ge & Zhang, 2011). The third reason for choosing the case was its institutional environment. One of Qingdao's planning 110 goals is to manage resources for the benefit of citizens and the ecosystems on which the city 111 depends. Qingdao and the Jiaozhou Bay play a key role in the first national-level marine 112 economy development strategy, paving the way for Shandong Province to be in the forefront 113 of coastal planning and management in China. As such, there are comprehensive rules about 114 coastal ecological protection in Jiaozhou Bay area, giving rise to a promising institutional 115 context for many related strategic plans (e.g. Qingdao Provisions of Marine Environment 116 Protection). These existing strategic plans attempt to address ES conflict issues by redefining 117 spatial use and managing activities to ensure local sustainable development. 118

119 3. A four-step method to analyze ES interrelationships

In general, coastal strategic planning for Jiaozhou Bay features activities for exploiting, utilizing and protecting coastal and marine resources. However, the impacts and extent of these activities on a set of ESs vary considerably since ESs are inevitably interconnected. We used a four-step method to investigate how activities, trade-offs and synergies among ESs were portrayed in coastal strategic plans. Meanwhile, reading the plans systematically enabled us to understand how plans are organized under a broad institutional environment, and to understand institutional implications to improve the inclusion of ES interrelationships.

127 Step 1: Selecting strategic plans

We focused on strategic spatial plans formulated during the last five years and collected four 128 strategic plans for Jiaozhou Bay from official websites and the responsible authorities (Table 129 1). The "Conservation and Development around Jiaozhou Bay" Strategy of Qingdao (Plan 1) 130 in 2008 was the first of these plans to promote the concept of integrating ecological protection 131 132 with industrial development for Qingdao City. It was an important urban space development strategy that enabled Qingdao to be part of The Development Plan of Shandong Peninsula 133 Blue Economic Zone (Plan 2). This plan is the first national sustainable development strategy 134 with a marine economy theme that highlights optimizing both seascape and landscape, 135 producing modern marine industrial systems and enhancing marine ecological civilization. 136 137 Two statutory urban strategic plans - The Twelfth Five-Year National Economic and Social Development Plans of Qingdao (Plan 3) and The Overall Urban Plan of Qingdao (2011-2020) 138 139 (Plan 4) – also reflect the role of coastal and marine resources in Jiaozhou Bay in improving

- 140 citizens' well-being and the urban economy. Gaining insight into which and how activities
- 141 and ES interrelationships may be integrated into these strategic plans can enhance the
- 142 adaptivity and sustainability in urban, regional and even national development.

Overall, given the emphasis these strategic plans place on interrelationships between ESs delivered by Jiaozhou Bay and regional/local development, we assumed that these plans have to address issues such as the organization, protection and development of activities that impact multiple ESs. Furthermore, as these are all strategic-level plans, they include a whole range of coastal activities. This could be useful for identifying more ES interrelationships caused by all these activities that are commonly found in coastal areas.

149 Step 2: Identifying ESs

150 Our previous study already identified the coastal ESs included in the four strategic plans' efforts (Li et al., 2015). We used a content analysis method accompanied by text 151 interpretation. To ensure coding consistency, a ES coding system was established based on 152 the four standard classification system put forward in the Millennium Ecosystem Assessment 153 154 (MA, 2005), which was complemented with other research particularly focused on coastal and marine ESs. There were several reasons for choosing the MA classification. First, the four 155 categories play a fundamental role because other modified classification schemes have widely 156 employed them as a foundation (e.g. Haines-Young & Potschin, 2010; Atkins et al., 2011). 157 158 Second, in order to qualitatively identify how activities and ES interrelationships may be portrayed in strategic planning, it is appropriate to adopt the MA typology which has been 159 used as a basis for prompting the discussion of social preference and values towards the 160 environment (Bryan et al., 2010). This classification would thus serve our research goals 161 162 better than others, which aim at valuing ESs (Haines-Young & Potschin, 2010; Atkins et al., 163 2011), uncovering the processes of delivering benefits (De Groot et al., 2002; Wallace, 2007), 164 analyzing spatial characteristics (Costanza, 2008), and distinguishing between ES excludability and rivalness (Fisher et al., 2009). A third reason concerns the supporting 165 services. Current studies usually exclude supporting services or subsume them in the group of 166 regulating services to avoid double counting of ES values. However, in our case, double 167 168 counting should not be an issue since no values would be aggregated. In our method, it is 169 important to consider supporting services and their institutional environment because some supporting services (e.g. habitat protection, biodiversity and resilience maintenance) have 170 become popular in political discourses across the world. Fourth, to gain a broad view of how 171 coastal and marine resources are used and affected by human activities through strategic 172 planning, some important and traditional abiotic services (regardless of ecological production 173 processes), such as space for navigation, industrial development and infrastructure and 174

offshore wind, were added to the provisioning group as some authors have done, for example,Atkins et al. (2011).

Subsequently, we examined each selected strategic plan sentence by sentence in order to 177 178 identify each coastal ES listed in the coding system. If a type of ES was referred to in a way 179 that linked it to the meaning of an ES concept or that contains any example stated in the 180 coding system, it was marked (Li et al., 2015). We coded terms and phrases in the documents 181 by using manuscript extraction techniques and NVivo software. A range of well-established coastal ESs integrated in documents was accordingly identified (listed in Table 2). In this step, 182 all the references to ESs were noted, which permitted us to further analyze the ES 183 interrelationships as formulated by planners and policy-makers in the strategic plans. 184

185 Step 3: Identifying drivers, ESs and their effects

186 We identified the activities that act as drivers affecting the delivery of ESs, as well as the ESs 187 themselves. This analysis was based on an interpretation of narratives mentioning at least one activity and two coastal services as coded earlier. The different types of activities (i.e., key 188 189 drivers) that were highlighted and associated with certain ESs in these four plans were summed up in a table. Each of these mainly perceived relations was regarded and named as 190 one type. This allowed us to not only identify the main drivers, but also to consider more ESs 191 in this stage. The effects of these activities were analyzed according to two types of 192 193 mechanisms identified by Bennett et al. (2009): "effects of drivers on multiple ESs" and "interactions among ESs." Thus, the direction of the effect is either from drivers to ESs or 194 195 from ES to ES, that is, bidirectional or unidirectional. This can be interpreted through the 196 contents involving both the driver and ESs identified earlier. We considered words such as "cancel," "forbidden," "limit," "control," "reduce", or "avoid" as negative effects. Narratives 197 that included words such as "enhance," "stimulate," "provide," "explore," "preserve," 198 "restore," "create," "improve," "benefit", and "guarantee" were seen as indicating positive 199 200 effects, depending on their textual position.

201 Step 4: Constructing relational diagrams

We depicted the identified relationships in diagrams, providing a straightforward way to analyze the initial inclusion of activities, ESs involved and their effects as stated in the strategic plans. We employed the structuring method proposed by Bennett et al. (2009). In each relational diagram, the topmost rectangle is the driver affecting ESs and the rectangles below are ESs; the solid arrow indicates a positive influence, while the dotted arrow indicates a negative effect; arrows illustrate the directions of effects. We classified these relational diagrams in terms of trade-off and synergy. The former group focused on managing services 209 that may co-vary negatively (more of one means less of another; Ring et al., 2010), while the 210 latter group co-varies positively (more of one means more of another; Ring et al., 2010) as a result of certain activities. Each group was further classified in terms of the attributes of a 211 212 driver (i.e., shared or independent effects on multiple ESs) and the degree of ES interactions 213 (generally, the more ESs involved, the stronger the interactions would be). This step portrayed the relationships in a visual way, enabling us to observe which links were included 214 and which were overlooked. To confirm and complement the document-based analysis, we 215 216 then double-checked our assumptions by interviewing eight planners and policy-makers from key sectors who had been involved in any of the four plans. Key stakeholders for interviews 217 were mainly selected from six main institutions including the Shandong Peninsula Blue 218 219 Economic Zone Construction Office, the Shandong Environmental Planning and Design 220 Institute, the Qingdao Urban Planning Bureau, the Qingdao Ocean and Fishery Bureau, the Qingdao Environmental Protection Bureau, and the Qingdao Institute of Marine Geology. 221

222 4. Analyzing ES interrelationships in the strategic plans for Jiaozhou Bay

223 4.1 Inclusion of drivers and ESs

224 The Jiaozhou Bay strategic plans show attempts to concisely consider some relationships in 225 terms of trade-offs and synergies among coastal ESs that are impacted by human activities. 226 Table 3 summarizes the results, showing drivers and ESs identified through the second step of 227 content analysis across the four selected strategic plans. We found that various activities were 228 listed in plans, which in reality may influence ESs in different ways. However, there were ten 229 typical types (four trade-offs and six synergies) that could be mainly derived from the narratives of affecting ESs. Among all the activities identified in the four plans, three 230 231 (controlling reclamation, restoring natural shoreline, and building wetlands park/reserve) were referred to in all the plans. Plans 1 and 3 underlined two activities (i.e., constructing new town 232 and upgrading port function) for stimulating multiple ESs. The rest of the drivers were each 233 referred to at least once in at least one strategic plan. The "category" columns in Table 3 show 234 235 which category each service involved belongs to; this was done to facilitate a general awareness that the provisioning services were most often regarded to be under direct management. Cultural services more often appeared as positively co-varying services with 237 238 other ESs where synergies were concerned. The diagrams in Sections 4.2 and 4.3 reveal the 239 detailed interplay of driver-ES and ES-ES relationships as formulated and mentioned in these 240 strategic plans.

241 4.2 Trade-offs of ESs' inclusion

242 Figure 2 shows the four typical types of trade-offs that were considered and managed in the

243 four strategic plans for Jiaozhou Bay. Planners and policy-makers clearly recognized that 244 increasing some provisioning services can result in severe damage to other services. The 245 plans recommended various activities to directly limit certain provisioning services: for 246 example, "strengthen efforts to protect the coastline by stopping intertidal/pond aquaculture to 247 restore its natural coastal condition" (Plan 1, Type 3) and "designate island protected areas in which any economic development that may change the island's topography and 248 249 geomorphology is forbidden" (Plan 2, Type 1). The plans also referred to some (but not all) 250 indirect effects of coastal actions. For instance, Plan 4 (Type 4) acknowledged that "strictly 251 controlling the coastal development and construction projects around Jiaozhou Bay will limit the erosion of the bay area and water quality, thereby protecting the marine hydrodynamic 252 conditions and self-purification capacity"; meanwhile, it stipulated that industrial and port 253 254 businesses should not be allowed "to occupy high-quality beaches and shoreline" (Plan 4, Type 4). This suggests that the planners recognized the value of provisioning services in 255 influencing several regulating and cultural services. This kind of indirect influence can also 256 affect some supporting services (i.e., in Types 1 and 2) described in the four strategic plans. 257

Another driver-ES mechanism is a shared driving force that directly impacts multiple ESs rather than one. Although no specific references were given, the general knowledge and straight links between some certain drivers and ESs indicated that planners and policy-makers took them for granted. Here are two examples: 1) restoring the natural shoreline can directly create landscape value for cultural services (Type 3), and 2) defining an island's protected area can preserve natural conditions for biodiversity (Type 1).

264 4.3 Synergies of ESs' inclusion

265 Figure 3 illustrates the six typical types of synergies among ESs derived from the plans. 266 These synergies show that most of the drivers create direct and positive influences on 267 multiple ESs as a shared force in each relational type. The central focus of the drivers can be 268 categorized into two groups. The first group of drivers is related to ecological restoration 269 activities, such as establishing a wetlands park/reserve and restoring natural waterways (see Types 7 and 9). Drivers in this group directly stimulate cultural, supporting, and regulating 270 services. Plan 1 underlined several outcomes arising from the provision of an urban wetlands 271 272 park or reserve, including "moderately developing eco-tourism" and "enhancing the urban spatial landscape." Meanwhile, the benefits of wetlands park or reserve "restore the waterfowl 273 274 habitat to promote the conservation of wetland biodiversity and urban self-purification" 275 (Plans 1 and 3). The activity of restoring natural waterways (Type 7) was only discussed once in Plan 1: it was aimed at "creating a chain of ecological islands in northern Jiaozhou Bay," 276 "enhancing the capabilities of urban areas to prevent damage from flooding, drainage and 277

storm surges," and "increasing the environmental capacity for better water quality." In these cases, there were interrelationships between regulating and cultural services; relationships between supporting and regulating services were not described at all. Only two pairs of services, i.e., wetlands habitat and biodiversity maintenance, and wetlands habitat and tourism, were often cited together in all the documents studied, indicating bidirectional relationships. The other group of drivers concerns developing an integrated functional area. On the one

hand, these drivers can directly provide spatial and resource advantages for activities such as "creating a tourism industry that features a large industrial port" (Plan 1), "developing highefficiency agriculture in coastal areas within a leisure and tourism corridor" (Plans 2 and 4), and "establishing multi-functional urban areas with an exhibition business, a residential area, leisure activities, marine research and history based on the local ecological environment" (Plans 1 and 3). On the other hand, these examples contained no detailed information about how the wide range of ESs could be enhanced together or how they could produce negative effects.

292 **5. Discussion**

293 5.1 Reflection on the inclusion of ES interrelationships

The case study results demonstrate how the four-step method presented in this paper could be 294 295 useful in identifying a range of drivers and ES interrelationships implicitly considered by 296 planners and policy-makers. The results of the analysis will remind policy makers of the need to focus on intangible, vulnerable services and indirect impacts, which could contribute to 297 298 reducing conflicting uses and enhance the integration of interests in planning processes. Our findings suggest that planners and policy-makers in the Jiaozhou Bay case emphasize the 299 300 need to encourage certain coastal activities, which at the same time limits trade-offs of 301 different services, and constrains their synergies.

To put this understanding in a further international context, Table 4 illustrates a review of 302 303 international case studies on ES interrelationships derived from recent international literature. 304 These cases confirm that trade-off decisions, as perceived by decision-makers, experts, 305 researchers and communities, show a general preference for provisioning services. As 306 suggested by some scholars (Carpenter et al., 2006; Rodríguez et al., 2006; Hauck et al., 307 2013), two main reasons may explain why trade-offs are frequently linked to provisioning services. One could be that this group of services are utilized in regard of exclusive types of 308 spatial use (i.e. landscape or seascape), and another reason is that they are highly tangible and 309 always directly identified. Our findings accord with these general assumptions and reported 310 311 findings. However, in the Jiaozhou Bay case, there appears to be a relatively broader

312 consideration of the negative impacts caused by an emphasis on provisioning services:

313 management that sets sights on providing a single provisioning service will typically reduce

314 biodiversity and other services (Ring et al., 2010). Therefore, planners and policy-makers

315 have attempted to reduce or restrict such negative impacts by spatially locating and

316 developing strategies for ES provision.

317 Our findings are also in agreement with other research that found regulating services and 318 supporting services are more likely to shape synergistic links (Table 4). In Jiaozhou Bay, there was an increasing focus on conserving and restoring the supporting services (e.g., 319 wetlands habitat and biodiversity). Chinese planners and policy-makers have invested in 320 321 supporting services rather than solely in provisioning services, with the former aiming at generating multiple benefits and avoiding a tension between development and the 322 environment. However, the four plans failed to fully recognize many indirect effects of these 323 324 activities on other ESs created through supporting services. For instance, defining an island protection area (Type 1) could maintain the habitat function. The long-term maintenance of 325 326 coastal and marine habitats would increase biodiversity, which may provide an enormous 327 fishery resource from the reserve because of the spillover effect (Grafton & Kompas, 2005; 328 Shen et al., 2011). Moreover, maintaining the habitats may contribute to landscape protection 329 as well as cultural heritage, benefiting scientific research and education (Ma et al., 2013). 330 Interrelationships pertaining to regulating services were also generally underappreciated (e.g., 331 carbon storage, algal blooms prevention, and erosion and siltation control). The plans barely 332 reflected indirect contributions that natural regulating services would make to ecosystem 333 resilience and other services, which has been highlighted by researchers such as Bennett et al. 334 (2009). Reduced stress on natural services could result in an overemphasis on the engineered 335 infrastructure as well as the loss of coastal buffering and other regulating services (O'Farrell et al., 2012). Therefore, we argue that these partial and fragmented acknowledgments fail to 336 identify the bundle of ESs directly and indirectly affected by a driver, which likely results in 337 338 an unbalanced appreciation of different ES categories.

339 Similar to several cases researched by other scholars (Rodríguez et al., 2006; Halpern et al., 340 2008), the selected strategic plans put little emphasis on temporal and spatial issues that were crucial for ES interrelationships. In the governance of Jiaozhou Bay, planners and policy-341 makers mainly focused on provisioning services at the local scale (e.g., agriculture, transport 342 343 and navigation services). They overlooked the spatial aspect of regulating and supporting 344 services that, "although delivered at a local scale, are dependent on ecological functioning that span broader spatial boundaries" (Duraiappah et al., 2014). One example is the wetlands 345 346 park, which could be influenced by pollution from the upper reaches outside administrative 347 boundaries - its management plan was restricted to the local scale. The frequency of activities 348 relative to ecosystems' temporal dynamics is also critical for a better understanding of how a

349 particular activity influences ES changes (Halpern et al., 2008). However, only the

350 management of reclamation restriction in the bay indicated an awareness of the need to

351 control long-term severe cumulative impacts. There was no other mention of such awareness

- in the plans. Accordingly, this weakness may nullify the definition of acceptable levels of
- 353 activities permitted under certain ES levels, and affect decisions about how much one ES can

354 be sacrificed in order to obtain another (Halpern et al., 2008).

Overall, the outcomes reported give planners and policy-makers insights into the importance of using multiple ESs by managing their interrelationships at different temporal and spatial scales. However, it is also important to recognize that clarifying ES interrelationships is not a simple task in practice. Strategic planning and policy-making will also face new challenges: for instance, how ES interrelationships can be comprehensively interpreted, when it is necessary to broadly balance different ESs, and how governance can maintain a grip on ES trade-offs and synergies.

362 **5.2 Institutional implications**

363 Not only did our method reveal interrelationships among ESs pertaining to diverse activities 364 considered in coastal strategic plans but the method and the results also point out several 365 reasons to explain the different levels of inclusion of drivers and ESs in the strategic 366 documents of Jiaozhou Bay. These outputs could enhance actors' ability to reflect institutions 367 and governance systems that fundamentally determine drivers and ES interrelationships.

368 First, our results show that strategic planning mainly underlines coastal economic development activities (e.g. the construction of agriculture, new towns, regional industrial 369 370 cultural clusters and sea ports) to create multiple ES synergies associated with higher market 371 value rather than ecological importance. This emphasis is understandable due to the socioeconomic focus, and the initial market-oriented preferences of the majority of related 372 373 authorities, particularly the coordinating sector that was responsible for each plan (see Table 1). The narrow ecological goals of most authorities probably lead to a lower diversity of 374 375 drivers that may prevent ES trade-offs. Second, the financial appropriation discussed in the strategic documents also implies a lack of balance in the focus on ecological protection and 376 377 marine economic activities. Funds could therefore wield a significant influence on activities 378 that may benefit regulating and supporting services. Third, we cannot overlook the 379 implications of the essentials of planning institutions on the inclusion of drivers and ESs. The 380 essentials include the mutually related national, provincial and local legislations and 381 regulations, and the approved specific plans focusing on, for instance, coastline protection and 382 comprehensive river regulation. These current institutional arrangements (e.g. the Marine 383 Functional Zoning, the Qingdao Provisions of Marine Environment Protection, and the 384 Reclamation Control Line) mainly formulate the spatial features of most activities in order to 385 avoid conflicts in ES use (see Figure 2). The arrangements also suggest that abiotic benefits 386 are usually best recognized by local authorities as they are easy to integrate into planning processes (Piwowarczyk et al., 2013). Moreover, as regards the spatial and temporal 387 388 mismatches, without a regional ES benefit-sharing institution based on broad cooperation, 389 objectives, such as "realizing environment co-protection, industrial interaction and 390 information sharing" across administrative boundaries" (Plan 1), were less likely to be met. Technical support was limited or not formally enhanced to strengthen the analysis of spatially 391 and temporally accumulative effects on ESs. Project-oriented and regionally-oriented 392 393 environmental impact assessments have proven to be particularly difficult for identifying spatial and temporal issues in strategic plans (Partidario & Gomes, 2013). 394

Overall, the analysis shows that when discussions of drivers and ES interrelationships were 395 396 integrated in the plans, they were usually specific to policy concerns present in the 397 institutional context in which the plans were embedded. Consequently, the existing 398 institutional arrangements in Jiaozhou Bay should be adjusted. Efforts could be invested in 399 enhancing initial ecological-value preference among planning sectors, expanding the scope of 400 ecological goals and the investments of environmental projects, promoting coastal-related 401 legislation and specific urban ecological plans, providing ES benefit-sharing schemes based 402 on a broad participation of stakeholders, and strengthening technical planning support by 403 integrating ES concepts.

404 5.3 Methodological reflection

405 We have developed a methodological framework, i.e. a four-step method, for identifying and 406 analyzing which and how different activities and ES interrelationships may be included in 407 coastal strategic planning. Content analysis has helped to establish straightforward and 408 detailed qualitative insights. Its advantage is generally more pronounced when a contextual understanding is required to understand how institutional settings shape the use of ES concept 409 (Piwowarczyk et al., 2013). Analytical tools that can inform such contextual understanding 410 411 would enhance decision-making on ES trade-offs and synergies through planning processes (Wilkinson et al., 2013). The typology promoted by Bennett et al. (2009) provides a more 412 causal description of ES interrelationships than the modeling and mapping methods 413 (Lautenbach et al., 2010). By adopting this typology, our method provides a step towards an 414 415 explicit identification of a set of policy interventions (i.e. drivers) that may modify 416 relationships of services. Not only the scope of underlying driving forces could be expanded

417 and observed, but a whole range of ESs was taken into account through the coding system.

418 This expanding perspective enables more comprehensive discussions on specific driving

419 elements and impacts than other single-issue ways, encouraging stakeholders to

420 straightforwardly realize that most of their benefits from ESs are vulnerable due to their

421 activities. Although we used a broad and perhaps partly inexplicit ES definition and

422 classification promoted by the MA (2005) to create the coding system, its flexibility leaves

423 sufficient space for further detailed mechanism analysis and, more importantly, an

424 understanding among multiple stakeholders about ES concepts and classifications.

425 The scope of the findings suggests that our method and the other three existing groups of 426 approaches, i.e., mapping, modelling, and social-survey analysis, in particular the social-427 survey analysis, could cross-fertilize each other. Apart from the contextual information and the broad scopes informed by our method, its qualitative understanding about planners' and 428 429 policy-makers' ways of implicitly managing activities and ES interrelationships are likely to enhance non-scientific audiences' acceptance of ES quantification approaches (Kelble et al., 430 431 2013). In turn, the explicitness and accountability of quantitative information concerning each ES-interrelationship mechanism can be supplemented by spatial, biophysical, economic and 432 433 social-value data. In particular, specific winners and losers created by certain drivers could be 434 investigated through social methods, which in turn may complement the identification of indirect ES interrelationships that have been ignored in planning. Therefore, links can be clear 435 436 between drivers and the benefits that related stakeholders may gain or lose from ES changes. 437 The identification of these links provides a way of translating social values back into 438 management strategies or even abstract goals for ES governance, and ultimately creates space 439 for solutions.

440 Our method would be useful to promote the identification of ES interrelationships during the real-life planning processes, making decision making more rational and informed. For 441 442 instance, in the early stage of defining the goals and the scope of plans, our method could 443 assist planners to consider the balance in social-economic goals and ecological goals that 444 affect drivers and related ESs, and to analyze the spatial and temporal scales for managing ES 445 s. During the stage of designing actions to achieve the goals, the visualized causal description could make the current proposal explicit and understandable for actors, reminding planners 446 some underlying links that have been previously overlooked. This method could also be 447 448 helpful to select different options on ESs together with quantifying approaches in biophysical, economic and social-value terms. In the stage of planning revision and approval, assessment 449 and suggestions on managing key drivers and their indirect, cumulative impacts to reduce 450 451 conflicts could be put forward based on this method. Finally, the visualized causal description 452 could work as a monitoring approach when patterns of natural resource or use evolve,

453 requiring adaptive solutions.

454 Overall, our approach is only a preliminary step towards incorporating ES trade-offs and 455 synergies into coastal strategic planning, and there are challenges facing implementation. First, 456 different planning and policy contexts determine which and to what extents diverse ESs can be acknowledged and employed within a coastal area. This is a key precondition for 457 458 identifying the majority of potential ES interrelationships and the effects of activities. 459 However, unclear identification of each service in strategic plans would probably restrict the 460 analysis of their relationships. Second, a dominant activity (one with an intensive or frequent influence) co-exists with other activities that have relatively minor effects (Halpern et al., 461 462 2008). This fact adds complexity to ES interrelationships and the long-term cumulative impacts analysis. Thus, it is a real challenge to identify and manage all possible drivers and 463 the different extents of their impacts. Finally, given the guiding role played by strategic 464 465 planning, only a few detailed ES interrelationships could be described in these strategic documents. This issue suggests that a specific assessment focusing on explicit ES-interacting 466 467 analysis would be highly useful (e.g., as part of strategic environmental assessment, and 468 ecological assessments of landscapes). Moreover, quantifying ESs across landscapes or 469 seascapes and through time, and monitoring small changes in the relationships among 470 services is also difficult (Bennett et al., 2009), but it would further refine the approach.

471 6. Conclusion

This paper argued that a more explicit and integrated inclusion of trade-offs and synergies 472 473 among ESs will make coastal strategic planning more adaptive and sustainable, and that a 474 systematic method to identify and assess this inclusion is needed. We presented a four-step 475 research method that mainly depends on ES-interrelationship mechanisms to identify which 476 drivers and ES interrelationships may be formulated in policy language in coastal strategic 477 planning. Our approach revealed which driver-ES and ES-ES interrelationships (assessed in 478 terms of direct or indirect, and positive or negative impacts) should be included. Again, the 479 results showed that interrelationships involving regulating and supporting services were less appreciated in Jiaozhou Bay's strategic planning than those concerning provisioning and 480 481 cultural services, which is similar to most international case studies. The findings illustrated 482 several direct institutional implications for considering different drivers and ESs. The fourstep method used distinguishes itself among ES-interrelationship assessment approaches by 483 identifying a wide scope of drivers and ESs and their consequences based on a more causal 484 485 mechanism, broadening strategic planning discussions and making ES integration more explicit. Meanwhile, this methodology is valuable for reflecting the institutional context 486 487 underlying ES interrelationships, and for providing potential for quantitative measurements.

- 488 Lessons learned from more case analyses and scientific knowledge informed by multi-
- 489 disciplined research would benefit its further development. Although integrating ES
- 490 interrelationships into policy strategies is difficult, further efforts for developing ecosystem-
- 491 service thinking are appropriate, and will have to include efforts to invent policy rules for
- 492 fundamental services (regulating and supporting) and interactions between users and services.

493 References

- 494 Atkins, J.P., Burdon, D., Elliott, M., Gregory, A.J., 2011. Management of the marine environment:
- 495 Integrating ecosystem services and societal benefits with the DPSIR framework in a systems 496 approach. Marine Pollution Bulletin, 62(2), pp.215–26.
- Bennett, E.M., Peterson, G.D. and Gordon, L.J., 2009. Understanding relationships among
 multiple ecosystem services. Ecology Letters, 12(12), pp.1394-404.
- 499 Böhnke-Henrichs, A., Baulcomb, C., Koss, R., Hussain, S.S. and de Groot, R.S., 2013. Typology
- 500 and indicators of ecosystem services for marine spatial planning and management. Journal of 501 Environmental Management, 130 (30), pp.135-45.
- 502 Brown, K., Adger, W.N., Tompkins, E., Bacon, P., Shim, D. and Young, K., 2001. Trade-off 503 analysis for marine protected area management. Ecological Economics, 37(3), pp.417-34.
- 504 Bryan, B.A., Raymond, C.M., Crossman, N.D., Macdonald, D.H., 2010. Targeting the
- 505 management of ecosystem services based on social values: Where, what, and how?.
- 506 Landscape and Urban Planning, 97(2), pp.111-22.
- 507 Butler, J.R.A., Wong, G.Y., Metcalfe, D.J., Honzák, M., Pert, P.L., Rao, N., van Grieken, M.E.,
- 508 Lawson, T., Brucec, C., Kroon, F.J. and Brodiee, J.E., 2013. An analysis of trade-offs between
- 509 multiple ecosystem services and stakeholders linked to land use and water quality management in
- 510 the Great Barrier Reef, Australia. Agriculture, Ecosystems & Environment, 180(1), pp.176-91.
- 511 Busch, M., Gee, K., Burkhard, B., Lange, M. and Stelljes, N., 2011. Conceptualizing the link
- 512 between marine ecosystem services and human well-being: the case of offshore wind farming.
- 513 International Journal of Biodiversity Science, Ecosystem Services & Management, 7(3), pp.190-514 203.
- 515 Carpenter, S.R., Mooney, H.A., Agard, J., Capistrano, D., DeFries, R.S., Díaz, S., Dietz, T.,
- 516 Duraiappah, A.K., Oteng-Yeboah, A., Pereira, H.M., Perrings, C., Reid, W.V., Sarukhan, J.,
- 517 Scholes, R.J., Whyte, A. and Clark, W.C., 2009. Science for managing ecosystem services:
- 518 Beyond the Millennium Ecosystem Assessment. Proceedings of the National Academy of
- 519 Sciences 106(5), pp. 1305-12.
- 520 Carpenter, S.R., Bennett, E.M. and Peterson, G.D., 2006. Scenarios for ecosystem services: an 521 overview. Ecology and Society, 11(1), pp. 29.
- 522 Chisholm, R.A., 2010. Trade-offs between ecosystem services: water and carbon in a biodiversity 523 hotspot. Ecological Economics, 69 (10), pp.1973-87.
- 524 Costanza, R., 2008. Ecosystem services: multiple classification systems are needed. Biological 525 Conservation, 141(2), pp.350-52.
- 526 Costanza, R., d'Arge, R., de Groot, R., Farber, S., Grasso, M, Hannon, B, Limburg, K, Naeem, S,
- 527 O'Neill, R.V., Paruelo, J., Raskin, R.G., Sutton, P., van den Belt, M., 1998. The value of the
- 528 world's ecosystem services and natural capital. Ecological Economics, 25(1), pp.3-15
- 529 Crossman, N.D., Burkhard, B., Nedkov, S., Willemen, L., Petz, K., Palomo, I., Drakou, E.G.,
- 530 Martı'n-Lopez, B., McPhearson, T., Boyanova, K., Alkemade, R., Egoh, B., Dunbar, M.B., Maes,
- 531 J., 2013. A blueprint for mapping and modelling ecosystem services. Ecosystem Services, 4, pp.4-532 14.

- 533 De Groot, R.S., Wilson, M.A., Boumans, R.M., 2002. A typology for the classification,
- 534 description and valuation of ecosystem functions, goods and services. Ecological Economics,
- 535 41(3), pp.393-408.
- 536 Duraiappah, A.K., Asah, S.T., Brondizio, E.S., Kosoy, N., O'Farrell, P.J., Prieur-Richard, A.H.,
- 537 Prieur-Richard, A.H., Subramanian, S.M. and Takeuchi, K., 2014. Managing the mismatches to
- 538 provide ecosystem services for human well-being: a conceptual framework for understanding the
- 539 New Commons. Current Opinion in Environmental Sustainability, 7, pp.94-100.
- 540 Eigenbrod, F., Anderson, B.J., Armsworth, P.R., Heinemeyer, A., Jackson, S.F., Parnell, M.,
- 541 Thomas, C.D., Gaston, K.J., 2009. Ecosystem service benefits of contrasting conservation
- strategies in a human-dominated region. Proceedings of the Royal Society, B-BiologicalSciences, 276, pp.2903-11.
- Fisher, B., Turner, R.K., Morling, P., 2009. Defining and classifying ecosystem services fordecision making. Ecological Economics, 68(3), pp.643-53.
- 546 Ge, Y. and Zhan, J.Y., 2011. Analysis of the impact on ecosystem and environment of marine 547 reclamation-A case study in Jiaozhou Bay. Energy Procedia, 5, pp.105-11.
- 548 Gee, K., Burkhard, B., 2010. Cultural ecosystem services in the context of offshore wind
- 549 farming: a case study from the west coast of Schleswig-Holstein. Ecological Complexity, 7(3), 550 pp.349-58.
- 551 Haase, D., Schwarz, N., Strohbach, M., Kroll, F., Seppelt, R., 2012. Synergies, trade-offs, and
- 552 losses of ecosystem services in urban regions: An integrated multiscale framework applied to
- 553 the Leipzig-Halle region, Germany. Ecology and Society, 17(3), pp.22.
- 554 Halpern, B.S., McLeod, K.L., Rosenberg, A.A. and Crowder, L.B., 2008. Managing for
- cumulative impacts in ecosystem-based management through ocean zoning. Ocean & CoastalManagement, 51(3), pp. 203-11.
- 557 Haines-Young, R., Potschin, M., 2010. Proposal for a Common International Classification of
- 558 Ecosystem Goods and Services (CICES) for integrated environmental and economic accounting.
- 559 Copenhagen: European Environment Agency. Available at:
- 560 http://www.nottingham.ac.uk/cem/pdf/UNCEEA-5-7-Bk1.pdf.
- 561 Hauck, J., Görg, C., Varjopuro, R., Ratamäki, O., Jax, K., 2013. Benefits and limitations of the
- 562 ecosystem services concept in environmental policy and decision making: some stakeholder
- 563 perspectives. Environmental Science & Policy, 25, pp.13-21.
- Holling, C.S. and Meffe, G.K., 1996. Command and control and the pathology of natural resourcemanagement. Conservation Biology, 10(2), pp.328-37.
- 566 Holt, A.R., Godbold, J.A., White, P.C., Slater, A.M., Pereira, E.G., Solan, M., 2011.
- 567 Mismatches between legislative frameworks and benefits restrict the implementation of the
- Ecosystem Approach in coastal environments. Marine Ecology Progress Series, 434, pp.213-28.
- 570 Kelble, C.R., Loomis, D.K., Lovelace, S., Nuttle, W.K., Ortner, P.B., Fletcher, P., Cook, G.S.,
- 571 Lorenz, J.J. and Boyer, J.N., 2013. The EBM-DPSER conceptual model: integrating ecosystem 572 services into the DPSIR framework. PloS ONE, 8(8), pp. e70766.
- 573 Gao, G.D., Wang, X.H., Bao, X.W., 2014. Land reclamation and its impact on tidal dynamics in 574 Jiaozhou Bay, Qingdao, China. Estuarine, Coastal and Shelf Science. (in press).
- 575 Grafton, R.Q. and Kompas, T., 2005. Uncertainty and the active adaptive management of marine 576 reserves. Marine Policy, 29(5), pp. 471-9.
- 577 Lautenbach, S., Volk, M., Gruber, B., Dormann, C.F., Strauch, M., Seppelt, R., 2010. Quantifying
- 578 ecosystem service trade-offs. In International Environmental Modelling and Software Society
- 579 (iEMSs) 2010 International Congress on Environmental Modelling and Software Modelling for
- 580 Environment's Sake. Ottawa, Canada.

- 581 Lester, S.E., Costello, C., Halpern, B.S., Gaines, S.D., White, C. and Barth, J.A., 2013. Evaluating 582 tradeoffs among ecosystem services to inform marine spatial planning. Marine Policy 38, pp. 80-9.
- 583
- 584 Lester, S.E., McLeod, K.L., Tallis, H., Ruckelshaus, M., Halpern, B.S., Levin, P.S., Chavez, F.P.,
- 585 Pomeroy, C., McCay, B.J., Costello, C., Gaines, S.D., Mace, A.J., Barth, J.A., Fluharty, D.L. and
- 586 Parrish, J.K., 2010. Science in support of ecosystem-based management for the US West Coast
- 587 and beyond. Biological Conservation, 143(3), pp. 576-87.
- 588 Li, R., et al., 2015. Ocean & Coastal Management (self-reference).
- 589 Li, T.H., Li, W.K., Qian, Z.H., 2010. Variations in ecosystem service value in response to land use 590 changes in Shenzhen. Ecological Economics, 69(7), pp.1427-35.
- 591 Liu, Z., Wei, H., Liu, G., Zhang, J., 2004. Simulation of water exchange in Jiaozhou Bay by 592 average residence time approach. Estuarine, Coastal and Shelf Science, 61(1), pp.25-35.
- 593 Millennium Ecosystem Assessment (MA), 2005. Ecosystems and human well-being: biodiversity
- 594 synthesis. Island Press: Washington, D.C., USA. [online] URL:
- 595 http://www.millenniumassessment.org/documents/document.354.aspx.pdf [Accessed 1 February596 2014].
- 597 Ma, C., Zhang, X.C., Chen, W.P., Zhang, G.Y., Duan, H.H., Ju, M.T., Li, H.Y. and Yang. Z.H.,
- 598 2013. China's special marine protected area policy: Trade-off between economic development and 599 marine conservation. Ocean & Coastal Management, 76, pp.1-11.
- 600 Martinet, V. and Blanchard, F., 2009. Fishery externalities and biodiversity: Trade-offs between
- the viability of shrimp trawling and the conservation of Frigatebirds in French Guiana. EcologicalEconomics, 68(12), pp. 2960-68.
- 603 Martínez-Harms, M.J., and Balvanera, P., 2012. Methods for mapping ecosystem service supply: a
- review. International Journal of Biodiversity Science, Ecosystem Services & Management, 8(1-2),pp. 17-25.
- 606 Martín-López, B., Iniesta-Arandia, I., García-Llorente, M., Palomo, I., Casado-Arzuaga, I.,
- 607 García Del Amo, D., Gómez-Baggethun, E., Oteros-Rozas, E., Palacios-Agundez, I., Willaarts,
- 608 B., González, J.A., Santos-Martín, F., Onaindia, M., López-Santiago, C., Montes, C., 2012.
- 609 Uncovering ecosystem service bundles through social preferences. Plos ONE 7, e38970.
- 610 Nelson, E., Mondoza, G., Regetz, J., Polasky, S., Tallis, J., Cameron, D.R., Chan, K.M.A.,
- 611 Daily, G.C., Goldstein, J., Kareiva, P.M., Londsdorf, E., Naidoo, R., Ricketts, T.H., Shaw,
- 612 M.R., 2009. Modelling multiple ecosystems services, biodiversity conservation, commodity
- 613 production, and tradeoffs at landscape scale. Frontiers in Ecology and the Environment, 7 (1), 614 pp.4-11.
- 615 Nelson, E., Sander, H., Hawthorne, P., Conte, M., Ennaanay, D., Wolny, S., Manson, S. and
- 616 Polasky, S., 2010. Projecting global land-use change and its effect on ecosystem service provision 617 and biodiversity with simple models. PloS ONE, 5(12), pp. e14327.
- 618 O'Farrell, P.J., Anderson, P.M., Le Maitre, D.C. and Holmes, P.M., 2012. Insights and
- 619 opportunities offered by a rapid ecosystem service assessment in promoting a conservation agenda 620 in an urban biodiversity hotspot. Ecology & Society, 17(3), pp. 27.
- Partidario, M.R. and Gomes, R.C., 2013. Ecosystem services inclusive strategic environmental
 assessment. Environmental Impact Assessment Review, 40, pp.36-46.
- 623 Piwowarczyk, J., Kronenberg, J., Dereniowska, M.A., 2013. Marine ecosystem services in urban
- 624 areas: Do the strategic documents of Polish coastal municipalities reflect their importance?.
- 625 Landscape and Urban Planning, 109(1), pp.85-93.
- 626 Potts, T., Burdon, D., Jackson, E., Atkins, J., Saunders, J., Hastings, E., Langmead, O., 2014.
- 627 Do marine protected areas deliver flows of ecosystem services to support human welfare?.
- 628 Marine Policy, 44, pp.139-48.

- 629 Raudsepp-Hearne, C., Peterson, G.D. and Bennett, E.M., 2010. Ecosystem service bundles for
- 630 analyzing tradeoffs in diverse landscapes. Proceedings of the National Academy of Sciences,
- 631 107(11):5242-7.
- 632 Ring, I., Hansjürgens, B., Elmqvist, T., Wittmer, H. and Sukhdev, P., 2010. Challenges in framing
- 633 the economics of ecosystems and biodiversity: the TEEB initiative. Current Opinion in
- 634 Environmental Sustainability, 2(1), pp.15-26.
- 635 Rodríguez, J.P., Beard Jr, T.D., Bennett, E.M., Cumming, G.S., Cork, S., Agard, J., Dobson, A.P.
- 636 and Peterson G. D., 2006. Trade-offs across space, time, and ecosystem services. Ecology and 637 Society, 11(1), pp. 28.
- 638 Salzman, J., Thompson Jr, B.H., Daily, G.C., 2001. Protecting ecosystem services: Science, 639 economics, and law. Stanford Environmental Law Journal, 20, pp.309-32.
- 640 Shen, G.Y., Huang, L.F., Guo, F., Shi. B.Z. eds, 2011. Marine Ecology (the third edition). Science 641 Press: Beijing, China. pp. 340.
- 642 Shen, Z.L., 2001. Historical changes in nutrient structure and its influences on phytoplantkon 643 composition in Jiaozhou Bay. Estuarine, Coastal and Shelf Science, 52 (2), pp.211-24.
- 644 Sherrouse, B.C., Clement, J.M., Semmens, D.J., 2011. A GIS application for assessing, mapping, 645 and quantifying the social values of ecosystem services. Applied Geography, 31(2), pp.748-60.
- 646 Swallow, B.M., Sang, J.K., Nyabenge, M., Bundotich, D.K., Duraiappah, A.K., Yatich, T.B.,
- 647 2009. Tradeoffs, synergies and traps among ecosystem services in the Lake Victoria basin of
- 648 East Africa. Environmental Science & Policy, 12(4), pp.504-19.
- Turner, R.K., 2000. Integrating natural and socio-economic science in coastal management.Journal of Marine Systems, 25(3), pp. 447-60.
- 651 Turner, K.G., Odgaard, M.V., Bøcher, P.K., Dalgaard, T., Svenning, J.C., 2014. Bundling
- 652 ecosystem services in Denmark: Trade-offs and synergies in a cultural landscape. Landscape
- 653 and Urban Planning, 125, pp.89-104.
- 654 Van der Biest, K., D'Hondt, R., Jacobs, S., Landuyt, D., Staes, J., Goethals, P., Meire, P.,
- 2014. EBI: An index for delivery of ecosystem service bundles. Ecological Indicators, 37,pp.252-65.
- 657 Villa, F., Ceroni, M., Bagstad, K., Johnson, G., Krivov, S., 2009. ARIES (Artificial
- 658 Intelligence for Ecosystem Services): A new tool for ecosystem services assessment, planning,
- 659 and valuation. In 11Th annual BIOECON conference on economic instruments to enhance the
- 660 conservation and sustainable use of biodiversity, conference proceedings. Venice, Italy.
- Wallace, K.J., 2007. Classification of ecosystem services: problems and solutions. BiologicalConservation, 139(3), pp.235-46.
- 663 Wilkinson, C., Saarne, T., Peterson, G.D., Colding, J., 2013. Strategic spatial planning and the 664 ecosystem services concept-An historical exploration. Ecology and Society, 18(1), pp.37.
- Chao, S.J., Jiao, N.Z., Shen, Z.L., Wu, Y.L., 2005. Causes and consequences of changes in
- 666 nutrient structure in the Jiaozhou Bay. Journal of Integrative Plant Biology, 47(4), pp.396-410.

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Table 4 Common ES trade-offs and synergies of different types of ecosystems analyzed by diverse methods

| No. | Document | Year | Sponsoring organization ^a | Implementing organization ^b | Source |
|--------|---|------|--|--|---|
| Plan 1 | "Conservation and Development Around Jiaozhou Bay" Strategy of Qingdao | 2008 | Qingdao Municipal Government | Qingdao Urban Planning Bureau (QUPB) | http://upb.qing dao.gov.cn |
| Plan 2 | The Development Plan of Shandong Peninsula Blue Economic Zone | 2011 | National Development and Reform Commission, the People's Government of Shandong Province | Shandong Province Development & Reform Commission | http://www.sdl b.gov.cn |
| Plan 3 | The Twelfth Five- year National Economic and Social Development Plans of Qingdao | 2011 | Qingdao Municipal Government | Qingdao Development & Reform Commission | http://www.qd dpc.gov.cn/qdd pc/ |
| Plan 4 | The Overall Urban Plan of Qingdao (2011- 2020) | 2012 | Qingdao Municipal Government | QUPB | QUPB records office (paper documents) |

Table 1. Summary of four strategic plans related to Jiaozhou Bay

a: The municipal government, provincial government and some national ministries mainly take the responsibility for developing strategic plans with regard to managing behaviors of communities and individuals.

b: A particular sector was assigned as the coordinating body to implement a plan. The coordinating sector would be assisted by all the other related sectors, an expert advisory committee and the general public in terms of providing diverse ES information for decision-making that lies with the municipal or provisional government.

| Category | ES & Examples |
|--------------|---|
| Provisioning | Fish & seafood |
| | Energy production (biomass fuel, offshore oil and gas, wind, tide and wave power) |
| | Biochemical and pharmaceutical uses |
| | Transport and navigation (use of waterways for shipping) |
| | Coastal space for industrial development and infrastructure |
| | Residential and industrial water supply (abstraction of water for residential and |
| | industrial purposes) |
| | Urban ecological intervals (dividing different developing groups/function zones) |
| Regulating | Prevention of floods, storms, tsunamis and typhoons (protection by biogenic structures) |
| | Seawater intrusion |
| | Algal blooms |
| | Erosion and siltation control (maintenance of productive sediments, mitigating the |
| | effects of sea-level rise) |
| | Water purification and waste treatment |
| | Climate regulation (balance and maintenance of the atmosphere) |
| Cultural | Tourism and recreation (beach tourism, sunbathing, diving, windsurfing and kite- |
| | surfing, fishing, spas and wellness centers, bird-watching) |
| | Cognitive values (education and research arising from the marine environment, school |
| | excursions, monitoring global environmental change and indicators of ecosystem |
| | health, long-term environmental records) |
| | Aesthetic beauty (landscape) |
| | Cultural heritage and identity (value associated with the marine environment itself) |
| | Sea sports (competitive sailing, yacht races and other seawater competitions) |
| Supporting | Maintenance of biodiversity |
| | Maintenance of habitats |

Table 2. Coastal ESs identified in the four spatial plans for Jiaozhou Bay (Li et al., 2015)

| Туре | Driver | Service A | Cate* | Service B | Cate* | Service C | Cate* | Service D | Cate* | Service E | Cate * |
|---------|--|--|-------|---|-------|---|-------|--|-------|--|-----------|
| Trade-o | off | | | | | | | | | | |
| 1 | Defining an island protection zone | Economic development that changes topography and geomorphology | Р | Biodiversity | S | | | | | | |
| 2 | Development of estuarial wetlands | Modern manufacturing industry | Р | Wetlands | S | | | | | | |
| 3 | Natural shoreline restoration | Intertidal/pond aquaculture | Р | Coastal aesthetic sense and landscape | С | Water purification | R | | | | |
| 4 | Shoreline division for reclamation control, industrial development, petrochemical zone control | Land use for industry, agriculture, port development | Р | Environmental capacity within the bay, self- purification capacity | R | Landscape resource | С | | | | |
| Synerg | у | | | | | | | | | | |
| 5 | Special agriculture construction | Marine food supply | Р | Leisure and tourism | С | | | | | | |
| 6 | Upgrading port function | Shipping | Р | Port tourism | С | | | | | | |
| 7 | Excavating artificial river, restoring natural waterways | Protection from flood and storm surge | R | Water purification | R | The landscape of ecology island chain | С | | | | |
| 8 | Constructing regional industrial cultural clusters | Marine culture | С | Tourism | С | Technology | С | | | | |
| 9 | Building wetlands park or wetlands reserve | Habitat protection | S | Ecotourism | С | Biodiversity | S | Urban air and water purification | R | Urban spatial landscape | С |
| 10 | New town construction | House | Р | Tourism | С | Wetlands | S | Business | Р | Marine scientific research, history & culture | С |

Table 3. Drivers and ESs of trade-offs and synergies included in strategic planning for Jiaozhou Bay

* Category: P-provisioning service, R-regulating service, S-supporting service, C-cultural service

| Source | Type of ecosystems | Study areas | Drivers | Trade-offs (vs.) | Synergies (&) | Methodology |
|---|---------------------------------|-----------------------------------|--|--|--|--|
| Piwowarczyk et al., 2013 ^a | Coastal | Polish coastal municipalities | No specific | (P-C) ports and fishery vs. beaches recreation (S-C) biodiversity vs. leisure activities (C-C) tourism vs. landscape | | Content analysis |
| Wilkinson et al., 2013 ^b Salzman et al., 2001 c | Urban Watershed | Melbourne and Stockholm USA | Land use change Water management | (P-R) timber production vs. freshwater supply (P-S, P-R) agricultural food vs. soil erosion, flood protection and protection of species | (P-C) agriculture and forestry production & recreational services (R-R) watershed preservation & flood control | Content analysis |
| Hauck et al., 2013 ^d | Agriculture, forestry, water | Finland, Germany, and Poland | No specific | • (P-S, P-R) industrial forestry vs. biodiversity, erosion, natural flood protection, purification of groundwater and natural carbon sinks | (S-P, C-P) biodiversity and tourism & organic agriculture (R-R, R-S) flood protection & water purification, erosion prevention, climate regulation and biodiversity | Survey, interview, focus group discussion |
| Holt et al., 2011 ^e | Estuary wetland | UK | No specific | • (P-C, P-R, P-S) fishing and farming vs. recreation, algae and biodiversity maintenance | • (C-C) aesthetic enjoyment & natural heritage | Workshop, content analysis |
| Potts et al., 2014 ^f | Marine | UK | Marine Protected Areas management | | (S-C) species & cultural wellbeing and tourism/nature watching (S-S, S-R, S-P, S-C) habitats & supporting, regulating, provisioning and cultural services | Expert workshop |
| Busch et al., 2011 ^g | Coastal | Schleswig-Holstein, German | Offshore wind farm construction | • (P-C, P-S) offshore wind vs. recreation and habitat | • (P-R, P-P, P-C) renewable energy production & climate regulation, fishery and marine culture | Questionnaire, researchers workshop |
| Martín-López et al., 2012 ^h | Territorial | Spain, the Iberian Peninsula | No specific | • (P-R, P-C) provisioning vs. regulating and almost all cultural services | - | Questionnaire, statistical analysis |
| Butler et al., 2013 | Floodplain | Tully–Murray | No specific | • (P-R) food and fibre production vs. water | • (R-C) water quality & floodplain | Statistical |

Table 4 Common ES trade-offs and synergies of different types of ecosystems analyzed by diverse methods

| | | catchment, Australia | | quality | recreational and commercial fisheries | analysis |
|-------------------------------|---------------------------|-------------------------------------|--|--|--|-----------------------------------|
| Raudsepp-Hearne et al., 2010 | Pre-urban agricultural | Quebec, Canada | No specific | • (P-R, P-C) crop and pork production vs. both regulating and cultural services | | ArcGIS, ES proxies |
| Turner et al., 2014 | Territorial | Denmark | No specific | • (P-C, P-R) crop production vs. sense of place, carbon storage, and wetland water purification | (R-C) carbon storage & sense of place and nature appreciation (P-P) crop production & livestock production | ArcGIS, ES proxies |
| Nelson et al., 2009 | Watershed | Willamette Basin, Oregon | Land use change | • (P-R, P-S) agricultural crop products, timber harvest, and rural-residential housing vs. hydrological services, soil conservation, carbon sequestration, and biodiversity conservation | • (S-R, S-P, S-C) biodiversity conservation & other ES | InVEST |
| Eigenbrod et al., 2009 | Watershed | Lake Victoria Basin, East Africa | No specific | • (P-R) agricultural production vs. sediment control | | Biophysical models and GIS |
| Gee, K Burkhar, 2010 | Forrest | Jonkershoek Valley, South Africa | Afforestation | • (P-R) timber production vs. water supply | • (R-P) carbon sequestration & timber production | Ecological- economic model |
| Haase et al., 2012 | Rural-urban | Leipzig-Halle region, Germany | Soil sealing; brownfield restoration | (P-C) food supply vs. recreation potential (P-R) food supply vs. climate regulation (C-R) recreation vs. carbon storage | (S-C) bird species diversity & recreation (P-R) food supply & carbon storage (S-R) biodiversity potential & carbon storage | Biophysical models, mapping |
| Van der Biest et al., 2014 | Watershed | Grote Nete Basin, Belgium | No specific | (P-R) food production vs. climate regulation (P-R) wood production vs. climate regulation | | Model and mapping |

a, b, c: ES trade-offs and synergies perceived by decision-makers and planners

f, g: ES trade-offs and synergies perceived by experts or researchers

d, e, h: ES trade-offs and synergies perceived by stakeholders (e.g. fishers, NGOs, planners, sectoral workers and local communities)

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plans



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